

Determining Representative National Surface Water Chemical Concentrations for Risk Prioritization Risa R. Sayre^{1,2,3}, Marc Serre², R. Woodrow Setzer, John Wambaugh¹

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1. Introduction

Background: With thousands of chemicals in commerce and the environment, efficient tools are needed to support risk prioritization and evaluation.

Knowledge gap: Inconsistent data availability for concentrations in surface water to develop exposure estimates.

mg/kg BW/day Potential Hazard from in vitro with Reverse Toxicokinetics Potentia **Exposure Rate**

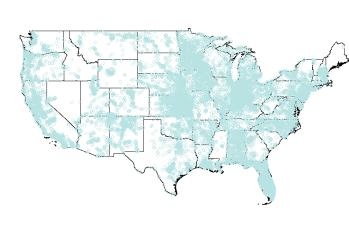
Proposed solution: Development of an open, reproducible workflow to:

- 1. Determine representative surface water concentrations for hundreds of organic chemicals in the United States based on already available monitoring data
- 2. Prioritize organic chemicals based on the relationship between concentration ranges and predicted no-effect concentrations (PNECs) for standard freshwater test species

2. Data overview and curation

The Water Quality Portal

(https://www.waterqualitydata.us/portal/) provided millions of concentrations of organic chemicals in surface water sampled from 2008 to 2018 covering broad spatial and physicochemical property ranges.



www.epa.gov/research

- 1626 names mapped to chemical structures using EPA's Chemicals Dashboard; 111 names manually curated
- 117 unmapped; 311 names referring to mixtures, ambiguous structures, organometallics manually removed

The final dataset contains 1404 unique structures.

Upper right: Sampling sites of observation set represent 2114 of 2270 hydrologic subbasins. Lower left: Chemical property space (log10, calculated using OPERA 2.4) of observation set: vapor pressure (mmHg), octanol:air, octanol:water, water solubility (mg/L).



3. Metadata filtering

Excluded sites:

- Not representative of ambient concentrations (Waste-injection well, sewer, finished water)
- Not surface water (Borehole, atmospheric)
- Not fresh water (Ocean, estuary)



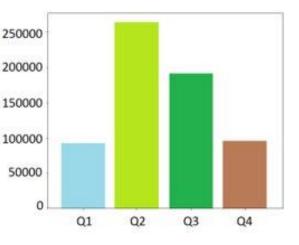


A palustrine wetland

- Surface (some edge cases like palustrine wetland, hyporheiczone/Ranney well, stormwater)
- Samples labeled as simply "water" without metadata
- Excluded activities: Not representative of ambient concentrations (blanks, spikes, leachate, initial dilution zone, radiolabeled)

4. Identifying representative subsets

Using two-sample Kolmogorov-Smirnov (KS) tests, we determine whether observed concentrations per chemical are "same" or "different", comparing sets by:



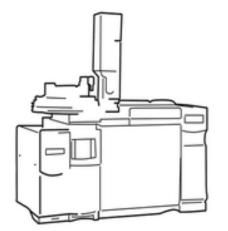
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Season

Although more samples were collected in warmer months, observations for about 90% of chemicals were not significantly different in magnitude

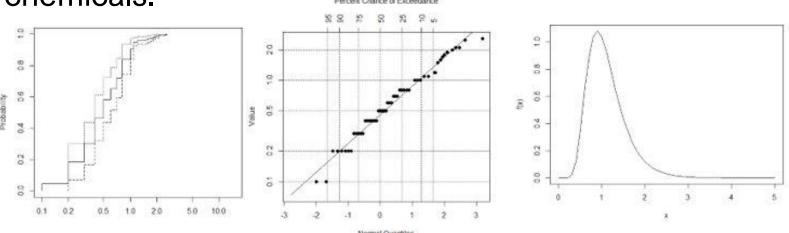
Sample phase

Determined from three metadata fields. For the 334 chemicals with both bulk and dissolved concentrations, 33% were significantly different.



Limit value type Observations above reporting limits, quantitation limits, and detection limits were **not** significantly different.

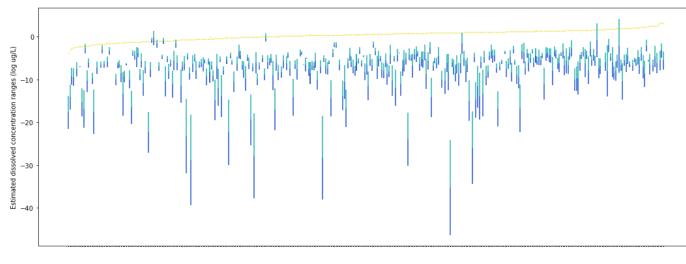
94 chemicals were excluded based on <2 observed values (measurements above the limit value) per sample phase. We evaluated three different methods of estimation by comparing the confidence interval (CI) around the mean concentration across 20 equal-sized censoring levels from <30% up to >99.5%. In every group, **MLE** had the smallest CI for the greatest number of chemicals. Concentratior



analysis.



358 dissolved estimated distributions were compared with PNECs based on the lowest of three TEST-predicted LC50s.



8. References

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5. Estimates of means using censored data

listributions using (left to right) Kaplan-Meier, robust regression on order (rROS), and Maximum Likellihood (MLE) for single chemical dissolved results with multiple censoring imits and 92.6% censored data

Chemicals with mean CI >1 μ g/L were excluded from further

7. Prioritization based on ecotoxicity estimates

The PNEC was within one standard deviation of the mean for nine chemicals and was not below any range.